

1 **Supplementary Information for “A bottom up approach to on-road CO₂ emissions**
2 **estimates: improved spatial accuracy and applications for regional planning.”**

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1 **Supporting Information**

2 All data used in this study are publically available, with sources as indicated in the references
3 section. Gridded results of our HPMS-based model are available for download at
4 <http://people.bu.edu/lrhutyra/Data.html> or from the corresponding author directly, as comma-
5 delimited files containing emissions estimates at 1km² and 0.1 degree grid scales.

6 All figures and spatial data were projected using the NAD1983 State Plane Massachusetts
7 Mainland FIPS 2001 Lambert Conformal Conic Projection.

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9 **1. Detailed Methodology of HPMS-based on-road CO₂ emissions model**

10 **Calculate VMT by functional class and county**

11 The HPMS is a national database managed by the Federal Highway Administration (FHWA)
12 that contains data on annual average daily traffic volumes (AADT) and centerline mileage for all
13 Federal-Aid roads and most other major and minor roads. We obtained annual VMT for each
14 road section in the HPMS by multiplying the average daily traffic volume by the length of the
15 road section in miles and then multiplied by 365 days. The FHWA requires that the AADT
16 values submitted by each state be adjusted prior to submission to take into account both
17 weekday/weekend and seasonal variations in traffic volumes on the road section. Thus the
18 AADT reported in the HPMS reflects an average daily traffic volume for any day in the calendar
19 year, independent of day of the week or month of the year. While this limits the use of the data
20 for analyses at shorter time scales, it allows for a straightforward estimation of annual statistics
21 for each road section without having to account for weekly and monthly variations.

22 The road sections in the HPMS are not geo-coded, and consequently we were not able to
23 assign our annual VMT estimates directly to a map of the road network. However, functional
24 class and county were available for each road section, with each functional class identified as
25 urban or rural depending on if the road section passes through a Census Bureau Urbanized

1 Area or Urban Cluster. Therefore we chose to aggregate our roadway-scale VMT to the county
2 scale, stratified by the 12 HPMS functional classes.¹ Since our roadway-scale HPMS data does
3 not include all of the VMT that occurred on local roads it was necessary to use a downscaling
4 approach to account for emissions from these roads. We allocated state-level data from FHWA²
5 on VMT for Rural Minor Collectors and Rural and Urban Local roads to each county using the
6 county's fraction of total state VMT as calculated from the HPMS dataset for each year.

7 **Disaggregate VMT by Vehicle Type**

8 As on-road CO₂ emissions are a product of fuel combustion, and the rate of emissions is a
9 function of fuel type,³ our intermediate goal was to estimate diesel and motor gasoline fuel
10 consumption for each functional class and county. First we partitioned annual vehicle miles
11 travelled amongst five different vehicle types: passenger cars, passenger trucks (which includes
12 SUVs, vans and pickup trucks), buses, single-unit trucks and combination trucks. State-level
13 data on the distribution of VMT among different vehicle types is available for the years 1993
14 through 1999 and for 2009 and 2010.⁴ A comparable national average distribution for the years
15 1980 to present exists as well.⁵ However, when we compared the state and national
16 distributions for 1993 through 1999 we observed that Massachusetts had significantly lower
17 fractions of passenger truck and heavy truck VMT across all road types relative to the national
18 average. Since this difference would strongly affect our fuel estimates, we chose to use the
19 state-level data for the available years and to estimate values for the years prior to 1993 and
20 after 1999. For our model years 1999 through 2008 we interpolated linearly between the state-
21 level distributions for 1999 and 2009; for years prior to 1993, we applied the 1993 distribution for
22 all years.

23 **Estimate fuel consumption by vehicle type, functional class, and county.**

24 We used the national average fuel economy for each vehicle type for each year⁵ to estimate
25 fuel consumption for each roadway functional class, county and year. Fuel consumption was
26 calculated as the quotient of distance travelled and average fuel economy. We assumed all fuel

1 consumption by passenger cars and passenger trucks was motor gasoline, all fuel consumption
2 by buses and combination trucks was diesel fuel, and that fuel consumption for single-unit
3 trucks was 23% motor gasoline and 77% diesel fuel. The fuel shares for single-unit trucks were
4 taken as an average value across the study period using reported fuel consumption by medium
5 and heavy vehicles obtained from the 2010 Transportation Energy Data Book.⁶

6 **Calculate CO₂ emissions by functional class and county**

7 We used emissions factors to estimate the CO₂ emissions produced by the fuel consumption
8 for each vehicle type. Fuel consumption was converted to CO₂ emissions using the emission
9 factors of 8.91 kg CO₂ per gallon gasoline and 10.15 kg CO₂ per gallon diesel fuel.³ We then
10 aggregated CO₂ emissions from both fuels to obtain total emissions for each functional class of
11 road at the county scale.

12 **Assign emissions to road network**

13 To assign emissions to a map of the road network we used the 2009 GIS Road Inventory
14 provided by the Massachusetts Department of Transportation⁷ which provides the length and
15 functional class of almost every road section in the state. We recognize that the road network
16 has changed in extent since 1980, but FHWA records for Massachusetts indicate that total
17 centerline mileage increased only 6.9%, from 33,777 in 1980 to 36,105 in 2008.⁸ We decided to
18 use the Road Inventory for our analysis, as it is the only geo-referenced dataset that covers all
19 Massachusetts roads. However, we note that the use of this dataset might introduce potential
20 errors due to the allocation of historical emissions across the contemporary road network.

21 To assign CO₂ emissions to each road we calculated the total centerline mileage of each
22 functional class of road in each county, and then divided our relevant CO₂ emissions by this
23 mileage to generate average per-mile CO₂ emissions. These average per-mile emissions were
24 then assigned by functional class and county to the road network for each year in the study
25 period.

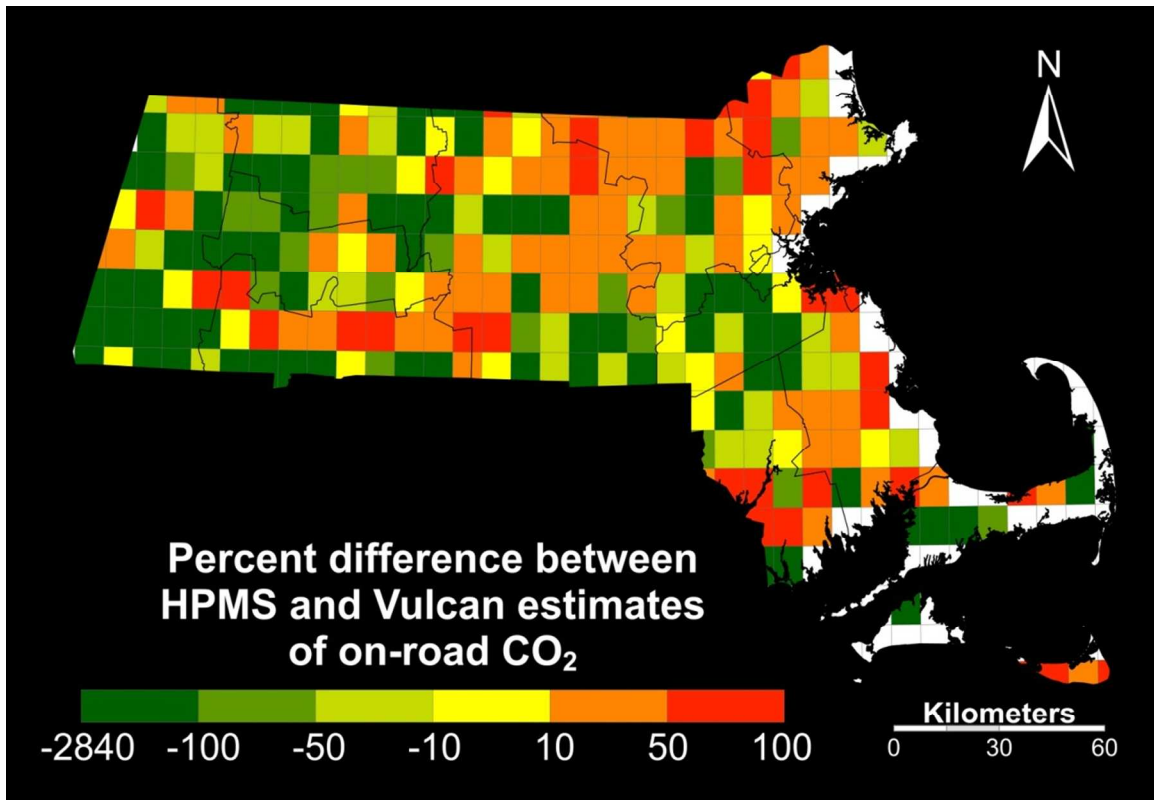
26 **Aggregate road-level emissions to other spatial scales**

1 For comparability with other CO₂ emissions inventories, we aggregated our roadway-scale
2 emissions to multiple scales: a 1km x 1km grid, a 0.1 x 0.1 degree grid, and summed to the
3 level of local towns. The 1km grid cells provided a high-resolution display of the emissions data
4 in a gridded format. The 0.1 degree grid aggregation provided for a direct comparison with the
5 EDGAR and Vulcan inventory products. The town level data was used in regression analysis of
6 the spatial and temporal correlations between emissions and population density, as population
7 data for the full time series was only available at the town scale. All figures and spatial data
8 were projected in ESRI ArcGIS 10, using the NAD1983 State Plane Massachusetts Mainland
9 FIPS 2001 Lambert Conformal Conic Projection.

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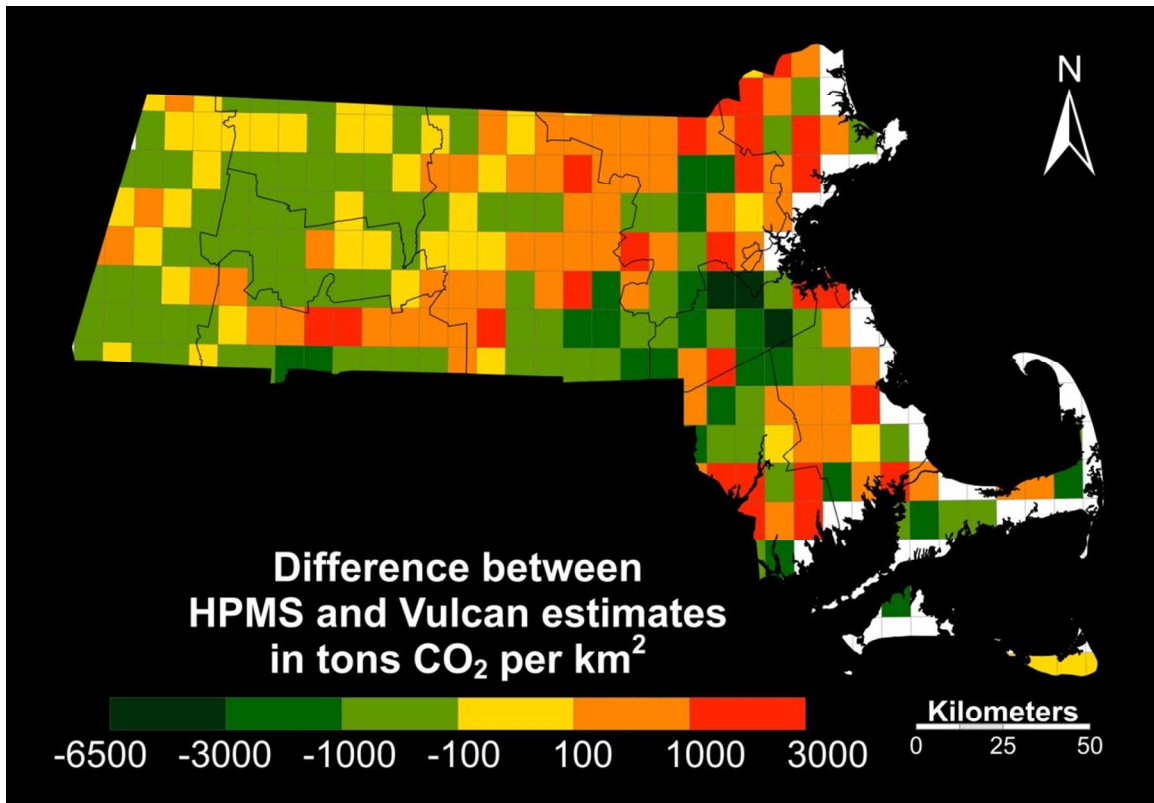
1 **2. Cell-by-cell comparison of HPMS and Vulcan emissions estimates for 2002**

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4 **Figure S1.** Percent differences between HPMS model CO₂ estimates and Vulcan Product
5 estimates for the year 2002, at 0.1 degree grid scale. Values calculated as: $(\text{HPMS} - \text{Vulcan}) /$
6 $\text{HPMS} * 100$; positive values indicate grid cells where HPMS estimates exceed Vulcan
7 estimates. White grid cells indicate cells where Vulcan reports zero emissions. This is a result of
8 the re-gridding process used by Vulcan to transform the original 10km² gridded results to the 0.1
9 degree grid.⁸ Note that the HPMS model produces higher emissions estimates than Vulcan in
10 urban areas, whereas the opposite is true in more rural or less populated areas.



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2 **Figure S2.** Absolute difference in estimates of tons CO₂ per square kilometer between HPMS
 3 and Vulcan. Positive values indicate cells where our HPMS model predicts higher CO₂
 4 emissions than the Vulcan Product. The spatial distribution of raw difference between the two
 5 models is similar to that of the percent difference, with HPMS producing higher estimates in
 6 urban areas relative to Vulcan (north central, north eastern and south eastern areas), and
 7 Vulcan producing higher emissions estimates in rural areas (western regions and parts of south
 8 central and south eastern areas).

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1 **Uncertainty Estimates**

2 The uncertainty associated with annual average daily traffic (AADT) values reported in
3 HPMS is characterized by the confidence intervals and precision levels reported in Chapter 6 of
4 the HPMS Field Manual.¹⁰ These confidence intervals vary by functional class, and take the
5 form of a combined confidence interval and precision level in the form of A-B, where A is the
6 probability that the value falls within B percent of the true value. For example a reported
7 uncertainty of 90-10 would mean that 90 percent of the time the reported value will be within
8 10% of the true value.

9 However, as each functional class of road has a different confidence interval and
10 precision level associated with its AADT estimates, these cannot be directly combined into an
11 overall estimate of the uncertainty in AADT for the whole HPMS data set. To standardize this
12 uncertainty, Mendoza et al.¹¹ converted each HPMS confidence interval / precision level into
13 one-sigma percent uncertainties as quoted below:

14 *“...the stated confidence interval and precision level were combined into a single estimate of*
15 *uncertainty as follows:*

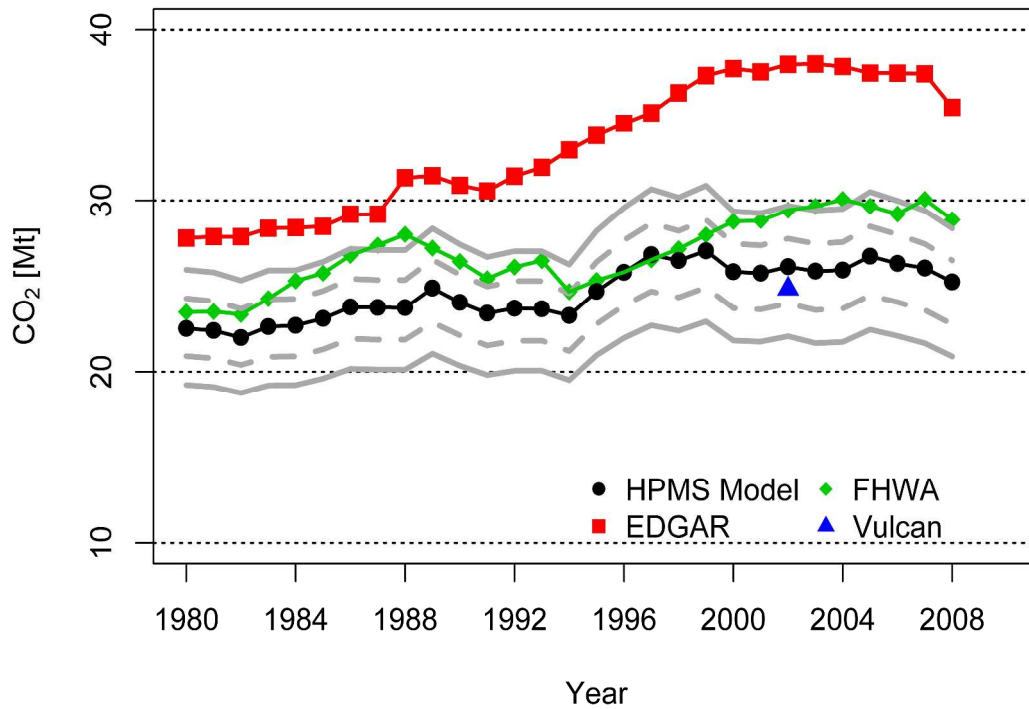
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$$U_x = V_x / S_x$$

17 *Where U_x is the uncertainty percent value associated with road type; V_x is the percent variation*
18 *from the true value for road type (10 for 90-10); and S_x is the number of standard deviations*
19 *within a normal distribution that is within variation of the true value for road type (“90” for 90-*
20 *10).”*

21 *Source: Mendoza et al.¹¹*

22 Using this method Mendoza et al. estimate one-sigma percent uncertainties for Rural Interstates
23 and Rural Principal Arterials of ±3.04%, for Rural Minor Arterials of ±6.08%, and for all other

1 functional classes of road of $\pm 7.8\%$. These are one standard deviation percent uncertainties,
2 which represent roughly a 68.3% confidence interval. To estimate a broader confidence interval
3 we also calculated a two-sigma uncertainty, equivalent to a 95.4% confidence interval. To do
4 this, we doubled the one-sigma percent uncertainties reported by Mendoza et al.¹¹ We used
5 both the one-sigma and two-sigma percent differences to calculate upper and lower estimates
6 of AADT for each road in our dataset. We then ran our model using these values to produce
7 upper and lower estimates of CO₂ emissions for each road section. The one-sigma shifted
8 AADT produced emissions estimates that ranged from $\pm 7.4\%$ to $\pm 7.6\%$ relative to our original
9 CO₂ estimates. The two-sigma shifted AADT produced emissions estimates that ranged from
10 $\pm 14.7\%$ to $\pm 15.2\%$ relative to our original estimates. Both of these ranges correspond well with
11 the empirical estimates of AADT uncertainty by Ritchie¹² and Gadda et al.¹³, and suggest that
12 the FHWA confidence/precision levels provide a reasonable basis for assessing the uncertainty
13 of CO₂ emission estimates generated from the AADT values reported by HPMS. We report the
14 upper and lower bound estimates of CO₂ emissions from our HPMS model in figure S3.



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2 **Figure S3.** Plot of CO₂ estimates from HPMS, FHWA, EDGAR and Vulcan. Solid gray line
3 shows upper and lower estimates from HPMS model run with two-sigma percent changes in
4 AADT. Dashed gray line shows upper and lower estimates from one-sigma percent change in
5 AADT.

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1 **Results of panel regression.**

Number of obs.	10150		
F(389, 9760)	1513.35	R-squared	0.933
Prob. > F	0.0000	Root MSE	104.29
Variable	Coefficient	Standard Error	t
pop / km ²	-26.684	2.971	-8.980
pop / km ² > 10	13.842	3.296	4.200
pop / km ² > 50	10.577	1.531	6.910
pop / km ² > 100	3.580	0.463	7.740
pop / km ² > 200	0.036	0.197	0.180
pop / km ² > 500	-0.418	0.100	-4.170
pop / km ² > 1000	-0.713	0.095	-7.530
pop / km ² > 2000	-0.346	0.082	-4.200
pop / km ² > 3000	0.130	0.060	2.170
pop / km ² > 4000	-0.243	0.073	-3.340
pop / km ² > 5000	0.140	0.073	1.900
pop / km ² > 6000	0.120	0.031	3.880
constant	775.924	85.122	9.120

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3 **Table S1.** Results of panel regression analysis. Dependent variable is tons of CO₂ emissions
 4 per mile of road. Town and year fixed effects not shown.

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6 **References for Supplemental Information**

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 8 Administration: Washington, DC, 2005; <http://www.fhwa.dot.gov/ohim/hpmsmanl/chapt4.cfm>

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